

矿山重金属污染土壤修复技术进展及展望

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内容提要:随着人们对矿山环境修复越来越重视,受矿山重金属污染土壤的修复技术不断向前发展,涌现出的新技术方法在矿山重金属污染土壤修复中发挥着日益重要的作用。本文通过对物理化学、植物、微生物和动物四大类矿山重金属污染土壤修复技术理论研究、试验和现场应用等方面入手,搜集大量资料,综述该四大类修复技术的研究现状和主要进展。总结提出矿山重金属污染土壤修复技术重点向4个方向发展:以低成本为导向的常规技术优化,包括常规廉价材料的有效利用、修复重金属污染的同时回收利用重金属(超积累植物和化学回收)等;以高精端新技术为导向的效率提升,包括纳米材料、生物薄膜等新型高效修复材料的研发、基因工程等,通过微观机理的精细研究大幅提高修复效率以降低总体成本;联合不同修复技术,如微生物—植物、化学—植物、物理—化学等联合修复技术,取长补短以实现更好的修复效果;加强不同修复技术的数据库和智能决策系统建设,促进技术成果转化。

关键词:重金属污染土壤;修复技术;植物修复;生物修复;纳米材料

矿产资源的开发在对国民经济发展起重要推动作用的同时,也带来了比较严峻的环境问题。矿山开采产生的废石、选矿产生的尾矿和冶炼废渣经风化淋滤等使有害元素转移到土壤中,造成土壤质量下降的同时污染农作物和地下水等,最后通过食物链进入人体,危害人类健康(赵元艺等, 2016)。世界上的多数国家也日益重视重金属对环境和人类健康的影响,并加入资金和技术投入进行矿山生态修复。以中国为例,我国大力倡导绿色矿山,积极推进矿山生态修复,据自然资源部发布《中国矿产资源报告(2019)》,2001年~2018年,累计恢复治理面积约 $100.46 \times 10^4 \text{ hm}^2$;其中,2018年全国新增矿山恢复治理面积约 $6.52 \times 10^4 \text{ hm}^2$,治理矿山7298个。

在矿山重金属污染研究中,重点关注的有害元素包括:Ni、Cu、Pb、Zn、Co、Au、Ag、Cd、As、Hg等。这些元素不同于有机质污染物,它们不会发生微生物或化学降解,因而会在土壤中停留相当长的时间(Adriano et al., 2004),不断积累,影响土壤性质。例如,Pb可以在土壤中持续存在150~5000 a

(Khalid et al., 2017), Cd的生物半衰期超过18 a (Förstner, 1995)。同时,土壤的物理化学性质,如pH、电导率、阳离子交换能力、土壤矿物学、微生物和生物学条件以及土壤无机和有机配体的存在,都会极大地影响土壤中重金属的活性(Laghlimi et al., 2015; Khalid et al., 2017)。因此,这就需要人们采用针对性的高效修复技术加快土壤修复进程。在过去十年,各国学者对矿山重金属污染修复进行了大量的研究工作,探索研究更为高效、经济的污染治理和修复技术,旨在降低环境中重金属的总量、生物活性以及它们在食物链中的积累(Bhargava et al., 2012; Sabir et al., 2015)。本文对国内外修复技术最新研究进展进行归纳,并对各修复技术进行了对比和总结,为我国矿山重金属污染土壤的修复研究提供理论依据和工作方法。

1 物理化学修复方法

1.1 物理方法

常见的重金属污染土壤的物理修复方法分为工

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工程措施法包括换土和深耕翻土法、客土稀释、分离修复法、隔离法等。客土和换土法主要分为深耕翻土、换土和客土。轻度污染土壤采用深耕翻土的方法,将表层含重金属离子土翻到底部,但不适用于矿山尾矿区;重污染土壤则采用异地客土的方法。这种方法对修复重金属污染土壤简单有效,但是工程量较大、投资高且容易造成土壤肥力下降等问题。因此这些方法通常适用于小面积的重金属污染土壤(Yao Zhitong et al., 2012)。此外,客土和换土法本质是将受污染土壤深埋,避免被植物根系吸收进入生物系统,但是受污染土壤中的重金属能够通过蒸散作用迁移至浅表土层,进而被植物吸收进入生物系统,因此如果种植农作物、经济作物或者是单纯的植被恢复,仍需要注意选择生物富集因子较小的植物。土壤隔离法是指采用防渗的隔离材料对重金属污染土壤进行隔离,主要应用于重金属污染严重,且难以治理的污染土壤。这种方法需要对隔离效果进行监测,以防止其他因素导致隔离失效。如 Gomez-Ros 等(2013)对 Cartagena—La Unión 采矿区物理修复进行了评估,发现采用 0.5 m 土壤进行隔离修复 30 a 后,顶层土壤中重金属的含量为 Cd(12 μg/g)、Pb(4616 μg/g)、As(67 μg/g) 和 Zn(3635 μg/g),相对应地区植物(黏蓬、山达树、百里香等)则含有高含量的 Cd 和 Zn,某种意义上而言,采用土壤隔离法不足以阻断底部与顶部之间重金属盐的对流,因此需要探索更有效和廉价的方法实现有效隔离。

热力修复技术则可细分为高温(约 1000 °C)原位加热修复技术、低温(约 100 °C)原位加热修复技术和原位电磁波加热技术等,主要针对具有挥发性的重金属污染物如汞(Hg)污染(郭维君等,2010)。

玻璃化技术是将受污染土壤加热到极高的温度(1700~2000°C)直至熔化然后快速冷却形成比较稳定的玻璃态物质(Navarro et al., 2013),这种玻璃态物质能够捕获并固定污染物,它具有低孔隙率和浸出活性,从而将其与环境隔离,但是其耗电成本较高。玻璃化技术结合采用一些新型改良剂如粉煤灰、活性炭和纳米金属等能够大幅度提高固定重金属污染物的效率(Mallampati et al., 2015)。

1.2 化学方法

重金属污染土壤化学修复方法主要可以分为三大类:

(1) 化学固定:是利用化学试剂将受污染土壤中的重金属污染物固定下来,降低重金属的生物有

效性;化学固定则是采用化学试剂如粉煤灰、水泥、二氧化硅和石灰等降低重金属的溶解度和活动性。重金属如 As、Pb 和 Cr 最适宜这种方法,其次是 Cd、Cu 和 Zn。Ok 等(2011)研究表明零价铁(zero-valent iron(ZVI),即还原铁粉)、石灰、腐殖质、堆肥以及这些化合物的组合能够有效固定 Cd 进而抑制种植大米对 Cd 的吸收,相比于未处理的参考土壤,经过改良剂处理的土壤中 Cd 的生物有效性降低约 50%~90%。此外,天然沸石也广泛用于受污染土壤中重金属(如 Cu、Co、Zn 和 Mn)的治理(Li Jie et al., 2018; Wen Jia et al., 2018; Boros-Lajszner et al., 2018)

(2) 加入化学试剂增强受污染土壤中重金属污染的活动性,将重金属污染物从土壤中去除,甚至回收,最典型的是化学浸出和离子交换。化学浸出主要是溶解重金属离子,进而将这些重金属转移进入浸出液,并将它们提取出来。浸出液通常是酸性的,以增加金属离子的溶解度。当重金属浓度较高且作用面积较大时优先采用化学浸出法(Alghanmi et al., 2015)。这种方法需要大量的酸以维持溶解重金属所需的 pH,后续还需要进一步中和酸化污泥,这也会大幅增加成本(Sharma et al., 2018)。

(3) 利用化学试剂将受污染土壤中的重金属污染物转化为低毒或无毒价态。这种方法主要是针对变价重金属污染土壤,应用相对比较局限。

1.3 纳米修复

近年来,具有优异的吸附能力、适度的稳定性和环保性能的新型纳米材料,已在捕获重金属离子方面取得了长足的发展,因此在这里单独介绍。应用于重金属污染环境修复的纳米材料包括金属有机框架(metal—organic frameworks(MOFs)),是一类由有机配体和金属离子或团簇通过配位键自组装形成的具有分子内孔隙的有机—无机杂化材料金属配合物,其结构具有长程有序性)、纳米零价铁(nanoscale zerovalent iron(nZVI),纳米尺寸的零价铁,具有强还原性和吸附性)、二维过渡金属碳化物、氮化物或碳氮化物(two-dimensional transition metal carbides and nitrides(MXenes),它是一种新型的二维过渡金属碳化物或氮化物,其化学式为 $M_{n+1}X_n$ ($n=1, 2, 3, M$ 为过渡金属元素, X 为碳或氮元素))、石墨相氮化碳($g\text{-C}_3\text{N}_4$)、纳米纤维素、 TiO_2 、 CeO_2 、 ZnO 、氧化铁等。

金属有机框架由于其孔隙度高、表面积大、易于加工且结构多样,与其他材料相比具有更强的吸附

能力 (Ricco et al., 2015; Yang Jichun and Yin Xuebo, 2017; Li Jie et al., 2018)。在最常见的条件下通过单羧酸改性剂合成的 UiO-66(一种锆基金属有机框架材料), 在 pH 2.0 时对砷酸盐吸附能力高达 303 mg/g(Wang Chenghong et al., 2015); 羧基修饰金属有机框架(UiO-66(Zr)-2COOH)(一种锆基金属有机框架材料)可以通过螯合作用选择性的高效去除 Cu²⁺(Zhang Yutian et al., 2015); 同时, MOF 与其他功能材料杂化能够增强实际应用, 如杂化 Fe₃O₄@MIL-100(Fe)(一种改性复合材料, 以四氧化三铁为核, 在其表面沉积金属中心离子和有机配体原位合成 MOF(Fe)而得到)能够有效去除 Cr⁶⁺(Yang Qingxiang et al., 2016)。

纳米材料另一个研究热点对象为纳米金属, 其中最具代表性的为纳米零价铁。由于成本低、比表面积大、表面结合位点丰富、反应速度快且无二次污染, 纳米零价铁已被用作高效且环保的吸附剂(Sheng Guodong et al., 2016; Pang Hongwei et al., 2018; Wu Yihan et al., 2019; Li Zhangtao et al., 2020)。但是, 纳米零价铁不稳定、易聚积、易氧化和次生污染等挑战, 因此需要在纳米零价铁中加入官能团或稳定剂克服这些缺陷, 如表面改性的纳米零价铁、多孔材料负载的纳米零价铁和无机黏土矿物负载的纳米零价铁(Liu Minghui et al., 2015; Fu Rongbing et al., 2015)。

2 植物修复技术

植物修复被认为是环境友好、极具吸引力、美观、非侵入性、节能和具有成本效益的清理低—中度重金属污染场地的技术。通过植物修复技术相对简单而且容易实现, 一方面可以减少水土流失, 另一方面可以降低重金属的迁移扩散, 且回收和处理富集重金属的植物较为容易; 植物修复技术可以协同其他常规的修复技术作为修复工程的最后一步, 故应用最多。此外, 植物修复技术还可以应用于有经济价值元素的提取, 如 Au、Ni 和 Tl 等。植物修复的劣势也很明显, 它没有常规方法那么高效和快速。Robinson 等(2015)提出了植物修复现实的时间框架, 即<25 a, 这就要求生物积累系数达到 10 以上才能够将土壤中金属浓度降低 50%。利用超积累植物修复重金属污染土壤主要有 4 种类型: 植物吸收、植物挥发、植物固定和植物转化(Mahar et al., 2016;

Khalid et al., 2017; Shah and Daverey, 2020)。

2.1 植物挥发

植物挥发是指植物将其吸收与积累的重金属元素转化为可挥发形态生物分子, 并通过蒸发作用释放进入大气(表 1);当然, 这也可能造成周围环境的次生污染。目前植物挥发修复重金属污染土壤研究较少, 且主要是针对重金属元素包括 Hg、As 和 Se, 如十字花科植物芥菜(*Brassica juncea*), 对土壤中 Se 的清除能力达到 40 g/hm²(Padmavathiamma and Li, 2007); 超积累植物蜈蚣凤尾蕨(*Pteris vittata*)能够将土壤中的 As 转化为挥发态 As 化合物(37% 亚砷酸盐和 63% 砷酸), 且能够在类似于亚热带的温室环境下将该植物从土壤中吸收的 As 以挥发的形式去除(约 90%)(Sakakibara et al., 2010)。Zhang Siyu 等(2013)发现海洋真核微藻 *Ostreococcus tauri* L. 能够将 As 甲基化并转化为挥发态, 且对 As³⁺ 转化为挥发态的量比 As⁵⁺ 高 5 倍。Guarino 等(2020)则详细研究了芦竹(*Arundo donax* L.)在植物生长促进细菌(*Stenotrophomonas maltophilia* sp. 和 *Agrobacterium* sp.)作用下对 As 的修复能力, 结果发现 75% 的 As 通过蒸腾作用以挥发态进入大气。

2.2 植物固定

植物固定是指利用植物降低土壤中重金属的生物可用度和活动性, 防止其进入地下水和食物链, 从而减少其对环境和人类健康的危害(Sylvain et al., 2016)。显然, 植物稳定并不会降低污染土壤中重金属的含量, 而是将它们限制在根际带附近(沉淀并积聚), 进而阻止它们发生迁移(Bolan et al., 2011; Ali et al., 2013; Abbas et al., 2016)。众多学者对植物固定技术进行了广泛的实验室和野外研究(表 2), 发现在众多植物中部分柳树是植物固定重

表 1 植物挥发修复重金属污染物

Table 1 Various plants reported for phytovolatilization and the metals contaminants removed by them

植物	污染物	资料来源
凤尾蕨(<i>Pteris vittata</i>)	As	Sakakibara et al., 2010
真核微藻(<i>Ostreococcus tauri</i> L.)	As	Zhang Siyu et al., 2013
北美鹅掌楸(<i>Liriodendron tulipifera</i>)	Hg	Greipsson, 2011
芦竹(<i>Arundo donax</i> L.)	As	Guarino et al., 2020
芦竹(<i>Arundo donax</i> L.)	As	Mirza et al., 2011
粉美人蕉(<i>Canna glauca</i> L.)、香芋(<i>Colocasia esculenta</i> L.)、纸莎草(<i>Cyperus papyrus</i> L.)、狭叶香蒲(<i>Typha angustifolia</i> L.)	As	Jomjun et al., 2010

金属的理想植物对象,因为柳树具有较高的重金属耐受性、强生长扩张能力以及强烈的蒸散作用(Sylvain et al., 2016)。更为重要的是,植物固定技术结合其他修复技术,如使用土壤微生物和有机改良剂,能够大幅增加其效率(表2)。以金属硫化物矿床开采导致的酸化重金属污染土壤治理为例,Yang Shengxiang等(2016)针对中国南部的大宝山多金属硫化物矿区的酸化土壤($\text{pH}<3$)采用石灰和鸡粪作为改良剂辅助的植物稳定进行修复,5种耐酸植物在6个月后成功生长并实现覆盖,前两年的数据表明这种方法有效地阻止了土壤酸化;在重金属植物稳定方面,所有5种植物都表现出根中重金属的含量高于枝叶,其中狗牙根(*Cynodon dactylon* L.)、匍枝毛连菜(*Panicum repens* L.)和类

芦(*Neyraudia reynaudiana* (Kunth.) Keng.)3种植物的重金属稳定能力尤为突出,以两年后植物根中的重金属含量为例:狗牙根(Zn:249 $\mu\text{g/g}$, Pb:186 $\mu\text{g/g}$, Cu:210 $\mu\text{g/g}$, Cd:1.2 $\mu\text{g/g}$),匍枝毛连菜(Zn:212 $\mu\text{g/g}$, Pb:188 $\mu\text{g/g}$, Cu:125 $\mu\text{g/g}$, Cd:2.29 $\mu\text{g/g}$),类芦(Zn:192 $\mu\text{g/g}$, Pb:239 $\mu\text{g/g}$, Cu:115 $\mu\text{g/g}$, Cd:2.79 $\mu\text{g/g}$)。

值得注意的是,近些年各国学者对生物炭修复技术进行了大量的研究,并取得了显著的成果。生物炭具有较高的表面积、孔隙率、可变电荷和官能团,因此向受污染土壤(或尾矿等)中加入可以增加土壤持水量、 pH 、阳离子交换量(CEC)、表面吸附量,盐基饱和度和农作物抗病性,进而增强刺激微生物多样性、土壤发芽率、土壤和农作物生产力、地上

表2 植物稳定修复重金属所采用的植物及相应的改良剂

Table 2 List of hyperaccumulator plant species for phytostabilization and corresponding amendments

植物	固定元素	改良剂	资料来源
柳树(<i>Salix viminalis</i> 和 <i>Salix purpurea</i>)	As、Sb、Pb	-	Sylvain et al., 2016
紫羊茅(<i>Festuca rubra</i>)	Zn、Cd	植物内生菌	Burges et al., 2016
紫羊茅(<i>Festuca rubra</i>)	Pb、Zn	牛粪、家禽粪便、造纸污泥与灰泥混合	Galende et al., 2014
狗牙根(<i>Cynodon dactylon</i> L.),匍枝毛连菜(<i>Panicum repens</i> L.),类芦(<i>Neyraudia reynaudiana</i> (Kunth.) Keng.),狼尾草(<i>Pennisetum purpureum</i> Schum.),大叶桉(<i>Eucalyptus robusta</i> Smith.)	Cu、Pb、Zn、Cd	石灰和鸡粪	Yang Shengxiang et al., 2016
芦苇(<i>Phragmites australis</i>),芦竹(<i>Arundo donax</i>)	Zn、Cu、Pb、As	Fe-WTR、MSW-C	Castaldi et al., 2018
芒(<i>Miscanthus sinensis</i>)	Cd、Cu、Pb、Zn	富铁改良剂	Lee et al., 2014
夹竹桃(<i>Nerium oleander</i>),白毛叶岩蔷薇(<i>Cistus albidus</i>),熏陆香树(<i>Pistacia lentiscus</i>)	As、Cd、Pb	碳酸钙和猪粪(ATS)	Parra et al., 2016
蒿柳(<i>Salix viminalis</i>),细弱剪股颖(<i>Agrostis capillaris</i>)	Cu	堆肥	Touceda-González et al., 2017
芒草(<i>Miscanthus sinensis</i>)	Cd、Cu、Pb、Zn、As	<i>Pseudomonas koreensis</i> AGB-1	Babu et al., 2015
黑麦草(<i>Lolium perenne</i> L.)	Pb、Cu、Ni、Cd	玉髓、石灰石、活性炭	Radziemska et al., 2017
蓖麻(<i>Ricinus communis</i> L.)	Ni	-	Adhikari and Kumar, 2011
酸角(<i>Tamarindus Indica</i>)	Cd	-	Udoka et al., 2014
刺苋(<i>Amaranthus spinosus</i> L.)	Cu、Pb、Cd	-	Chinmayee et al., 2012
龙葵(<i>Solanum nigrum</i> L.),菠菜(<i>Spinacia oleracea</i> L.)	Pb、Cu、Cd、Cr	-	Dinesh et al., 2014
芥菜(<i>Brassica juncea</i> L.)	Cu、Co、Ni	生物炭和堆肥混合	Rodríguez-Vila et al., 2014, 2015

生物量以及植被覆盖(Tang Jingchun et al., 2013; Kelly et al., 2014; Rodríguez-Vila et al., 2014, 2015; Anawar et al., 2015)。

植物固定方法适用于植物提取不太理想的地区,并且采用植物稳定技术需要检测土壤的物理化学等条件是否发生变化,即是否保持最佳植物稳定重金属的条件(Bolan et al., 2011),将转运到地上部分的重金属控制在最小范围。

2.3 植物提取

植物提取技术是利用超积累植物根系从土壤中吸取一种或多种重金属,并将其转移、贮存到植物组织中,如树叶、枝干等,通过收割去除土壤中重金属;

这些积累植物能够提取比正常植物高 50~500 倍的重金属(Baker and Brooks, 1989)。通常,地面以上植物组织能够积累>100 μg/g(干重) Cd,>1000 μg/g(干重) Ni、Cu、Pb 或者 10000 μg/g(干重) Zn 和 Mn 就称这些植物为超积累植物;在天然环境条件下植物叶子干燥后能够称之为超积累植物的重金属含量门限为 Se、Tl、Cd 为 100 μg/g; Cr、Co、Cu 为 300 μg/g; As、Pb、Ni 为 1000 μg/g; Mn 为 10000 μg/g; Zn 为 3000 μg/g; Au、Ag 为 1 μg/g(Mahar et al., 2016)。有许多植物能够积累 Ni、Cu、Pb、Cr、Co、Zn、Cd 等重金属,如柳树、象草(Napier grass)、伴矿景天(*Sedum plumbizincicola*)、甘蓝油菜(*Brassica napus*)、

表 3 植物提取修复重金属所采用的植物及相应的改良剂

Table 3 List of hyperaccumulator plant species for phytoextraction and corresponding amendments

植物	提取元素	改良剂	资料来源
粉叶蕨(<i>Pityrogramma calomelanos</i> L.),凤尾蕨(<i>Pteris vittata</i> L.)	As	—	Ha et al., 2019
伴矿景天(<i>Sedum plumbizincicola</i>)	Zn、Cd	内生细菌	Ma Ying et al., 2015
龙葵(<i>Solanum nigrum</i>)	Cd、Pb	聚天冬氨酸和液体氨基酸肥料	He Xiaoman et al., 2019
蜈蚣蕨(<i>Pteris vittata</i> L.)	As、Cd、Zn	—	Zeng Peng et al., 2019
象草(Napier grass)	Zn、Mn、Cu、Pb、Cd、Cr、As	—	Ma Chongjian et al., 2016
蕹菜(Water spinach)	Cd	混合营养嗜酸菌	Hao Xiaodong et al., 2019
千穗谷(<i>Amaranthus hypochondriacus</i>),青葙(<i>Celosia argentea</i>),龙葵(<i>Solanum nigrum</i>),商陆(<i>Phytolacca acinosa</i>),伴矿景天(<i>Sedum plumbizincicola</i>)	Cd	—	Huang Rong et al., 2020
柳树(<i>Salix viminalis</i> , <i>Salix schwerinii</i> , <i>Salix dasyclados</i>)	Cu、Ni、Zn	石灰和双膦酸盐	Salam et al., 2019
芥菜(<i>Brassica juncea</i> L.)	Cd、Pb	—	Gurajala et al., 2019
油菜(<i>Brassica napus</i>)	Co、Cr、Cd、Cu、Ni、Zn、Pb, As	生物炭	Gascó et al., 2019
天蓝遏蓝菜(<i>Noccaea caerulescens</i>)	Cd、Zn	—	Martínez-Alcalá et al., 2016
天蓝遏蓝菜(<i>Noccaea caerulescens</i>)	Cd、Zn	生物炭	Rees et al., 2020
东南景天(<i>Sedum alfredii</i> Hance)	Cu、Pb、Zn、Cd	非致病性尖孢镰刀菌(<i>Fusarium oxysporum</i>)	Zhang Xincheng et al., 2012
Corrigiola telephiifolia	As、Pb	—	García-Salgado et al., 2012
山菊头桔梗(<i>Jasione montana</i>)	Cd、Zn	—	García-Salgado et al., 2012
洋地黄(<i>Digitalis thapsi</i>)	As、Cd、Cu、Pb、Zn	—	García-Salgado et al., 2012
芸芥(<i>Eruca sativa</i>)	Cr	恶臭假单胞菌(<i>Pseudomonas putida</i>) (ATCC 39213)	Kamran et al., 2017

东南景天(*Sedum alfredii Hance*)等,大多数植物能够积累Cu、Ni、Zn,这些金属是植物提取修复的最佳目标(表3)。植物提取重金属领域的一个重要研究热点为重金属超积累植物的筛选,通常选择根系发达、且能够产生特定分泌物提高重金属生物有效性的植物。其次,植物修复技术通常需要土壤中的重金属离子呈可吸收的溶解态,为了提高植物对土壤中重金属的吸收能力,通常会加入一些有机络合剂来增加土壤中重金属的生物有效性。此外,针对具有经济价值元素的植物提取修复研究也取得了一定进展,并且已经成为一个重点研究领域——植物采矿(Phytomining)。现在所有国家的采矿操作必须确保环境的可持续发展,鉴于此,世界各国的研究人员试图将植物萃取发展为经济有效的采矿方式,目前研究程度较高的Ni和Au。以金为例,各国学者已针对金矿尾矿和金污染土壤进行大量实验研究(表4),但是植物采矿对基质金含量要求较高,基本上高于我国浅表金矿工业品位1 ng/g的标准,随着未来金价的回升以及技术的不断进步将在未来使金植物采矿成为一种经济可行的工艺技术(Wilson-Corral et al., 2012; Sheoran et al., 2013)。

目前,一个十分有价值的工作是金属超积累植物数据库的建设。目前各国学者对植物修复或者植物采矿等进行了大量的研究。但是这些关于超积累植物的信息是分散、无组织的。因此,迫切需要建立一个超积累植物数据库,该数据库应当包含各超积累植物的基本关键信息,如目标元素、生长环境以及改良剂等。国外一些机构或学者已进行了相关工作,如加拿大的PHYTOREM数据库、英国ECUS公司的METALS以及澳大利亚昆士兰大学矿山修复中心管理的超积累植物数据库等。据昆士兰大学矿山修复中心管理的超积累植物数

积累植物数据库免费公开的数据显示,超积累植物达到721种(Ni:523,Cu:53,Co:42,Cr:1,Mg:42,Zn:20,稀土元素:2,Se:41,Ta:2,Cd:7,As:5,Pb:8),并且一些植物能够积累多种金属元素(表5)(Reeves et al., 2018)。

2.4 植物转化

植物转化(或植物降解)修复重金属污染土壤是通过植物代谢过程或者是通过植物产生的化合物促进土壤中有毒重金属转变为低毒形态。早期的研究表明细菌和真菌等微生物能够促进将有毒重金属转化为低毒性(或无毒性)状态,如*Pseudomonas maltophilia*能够转化和沉淀不同有毒的金属离子,*Aspergillus niger*则能够将不溶无机金属化合物(如ZnO,Zn₃(PO₄)₂和Co₃(PO₄)₂)转化为有机不溶金属草酸盐(Sayer and Gadd, 1997; Blake II et al., 1993)。近些年来植物转化修复重金属污染土壤方面的研究较少。Caçador和Duarte(2015)发现盐生植物能够将盐沼中高毒性的Cr⁶⁺转化为低毒性的Cr³⁺。

2.5 挑战和前景

植物修复重金属污染土壤是一种十分具有吸引

表4 不同植物采矿试验植物金浓度统计

Table 4 Concentration of Au in plants from different phytomining trials

介质	植物	螯合剂	Au (μg/g)		资料来源
			植物	介质	
尾矿	孔颖草属(<i>Bothriochloa macra</i>)	NaCN	24	1.75	Piccinin et al., 2007
	芥菜(<i>Brassica juncea</i>)	KCN	30	0.64	Anderson et al., 2005
	芥菜(<i>Brassica juncea</i>)	NaCN	39	0.64	Anderson et al., 2005
	向日葵(<i>Helianthus annuus</i>)	NaCN	19	2.35	Wilson-Corral et al., 2011
	锯缘落地生根(<i>Kalanchoe serrata</i>)	(NH ₄) ₂ S ₂ O ₃	10	2.35	Wilson-Corral et al., 2011
	石茅(<i>Sorghum halepense</i>)	NaCN	24	2.35	Rodriguez-lopez et al., 2009
	白车轴草(<i>Trifolium repens</i>)	NaCN	27	1.75	Piccinin et al., 2007
	玉米(<i>Zea mays</i>)	NaCN	20	0.64	Anderson et al., 2005
硅砂	<i>Berkheya coddii</i>	KCN	97	5	Lamb et al., 2001
	油菜花(<i>Brassica campestris</i>)	NH ₄ SCN	304	3.8	Wilson-Corral et al., 2012
	芥菜(<i>Brassica juncea</i>)	NH ₄ SCN	57	5	Anderson et al., 1999
	芥菜(<i>Brassica juncea</i>)	KCN	326	5	Lamb et al., 2001
	胡萝卜(<i>Daucus carota</i>)	(NH ₄) ₂ S ₂ O ₃	89	3.8	Msuya et al., 2000
	芹菜(<i>Rapahanus sativus</i>)	NH ₄ SCN	113	3.8	Msuya et al., 2000
土壤	沙漠蕨(<i>Chilopsis linearis</i>)	CH ₄ N ₂ S	296	5	Rodriguez, 2006
	沙漠蕨(<i>Chilopsis linearis</i>)	NH ₄ SCN	197	5	Rodriguez, 2006

力的方法,但是仍然面临一定的挑战:

(1)超积累植物较低的生物质和较慢的生长速率限制了植物提取效率,导致修复周期通常较长。

(2)受气候影响的热带和亚热带地区由于害虫和疾病影响导致一些植物的修复能力受到影响,甚至失效,同时必须避免引入入侵物种作为超积累植物,这样可能会影响本地植物的多样性。

(3)多数污染场地的土壤质地差、有机质含量低,造成植物难以生长或长势缓慢,导致植物修复效率降低,因此要选择抗逆性强的植物。

(4)很难移动那些紧密结合的那部分金属离子,即土壤中这些重金属污染物的生物有效性有限。

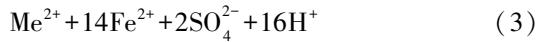
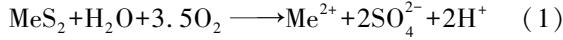
(5)需要在植物修复后正确处理这些受污染的生物质,错误处理生物质可能导致积累的金属转移进入食物链(Mahar et al., 2016; Ashraf et al., 2019)。

未来研究的重点在于不同植物作用机理、基因改造植物在植物修复中发挥的作用、联合采用多种植物应用于重金属污染土壤修复、数据库的建设以及后续效率评价等(Parmar and Singh, 2015; Nwaichi and Dhankher, 2016; Reeves et al., 2018)。

3 微生物修复技术

3.1 生物淋滤

生物淋滤技术是通过微生物将重金属或其化合物氧化或分解,将重金属呈溶解态释放,再回收的方法。其机理可以分为两种:一种是直接吸附在金属盐表面与重金属化合物发生反应[方程式(1)];另一种是间接细菌氧化,如铁氧化细菌能够形成 Fe^{3+} ,然后 Fe^{3+} 再氧化分解重金属化合物形成溶解态重金属[方程式(2)、(3)]。



目前这方面的研究主要集中在贵金属的生物浸出。近些年各国学者对金的生物淋滤进行了大量的研究,并取得了一系列进展,实验室条件下个别案例

表 5 全球数据库(Global Database)统计的已报道的最高浓度超积累植物种类
(Reeves et al., 2018)

Table 5 Hyperaccumulator species in the Global Database (as of September 2017) with the global records that are the highest concentrations reported to date(Reeves et al., 2018)

元素	门限值 ($\mu\text{g/g}$)	科	属	种	全球记录
As	>1000	1	2	5	蜈蚣蕨(<i>Pteris vittata</i>) (2.3%)
Cd	>100	6	7	7	叶芽鼠耳芥(<i>Arabidopsis halleri</i>) (0.36%)
Cu	>300	20	43	53	<i>Aeolanthus biformifolius</i> (1.4%)
Co	>300	18	34	42	星香草(<i>Haumaniastrum robertii</i>) (1%)
Mn	>10000	15	24	42	新喀里多尼亚维罗提(<i>Virotilia neurophylla</i>) (5.5%)
Ni	>1000	52	130	532	<i>Berkheya coddi</i> (7.6%)
Pb	>1000	6	8	8	<i>Cepaeifolia</i> (0.8%)
La,Ce	>1000	2	2	2	铁芒萁(<i>Dicranopteris linearis</i>) (0.7%)
Se	>100	7	15	41	黄芪(<i>Astragalus bisulcatus</i>) (1.5%)
Tl	>100	1	2	2	李果荠(<i>Biscutella laevigata</i>) (1.9%)
Zn	>3000	9	12	20	天蓝遏蓝菜(<i>Noccaea caerulea</i>) (5.4%)

中金采收率可高达90%以上,甚至100%(表6)。

3.2 生物稳定

生物稳定法则是将不活动的细菌作为增溶剂,将溶解态重金属转化为更加稳定的结合态,降低其活性和生物利用度。Li Xin等(2017)采用聚乙烯醇(PVA)和固定化硫酸盐还原细菌(SRB)将受污染沉积物中的Cu、Zn、Pb和Cd降低76.3%、95.6%、100%和91.2%。

3.3 生物薄膜

生物薄膜是一种高效的生物修复工具和生物稳定剂。生物薄膜修复环境机理包括生物还原、生物吸附、生物沉淀、生物络合、生物螯合和生物积累。可以归纳为两种途径:一是作为吸附剂等将重金属固定下来,降低其活性;二是细胞膜中存在的胞外聚合物质包含一些分子具有表面活性剂或乳化剂性质能够增强污染物的生物利用度。

生物膜通过吸附作用修复重金属污染是最常见的,如微藻生物膜PSBR通过吸附作用去除矿山垃圾渗滤液中的Zn(Li Tong et al., 2015),胶红酵母(*Rhodotorula mucilaginosa*)生物膜通过吸附作用去除 Hg^{2+} 、 Cu^{2+} 和 Pb^{2+} 等重金属(Grujic et al., 2017),Wu Gang等(2010)则提出一种自然生物膜能够去除极酸性和中等pH值得土壤渗滤液中的 Cu^{2+} 和 Cd^{2+} 。值得注意的是,生物膜的修复效率通常较高,

如胶红酵母(*Rhodotorula mucilaginosa*)以浮游细胞的形式金属去除效率为4.79%~10.25%,而以生物薄膜的形式则为91.71%~95.39%(Grujic et al., 2017)。

此外,针对含金属硫化物的钙质尾矿修复,García-Meza等(2011)提出了一种光合生物膜,这种生物膜能够避免和减少尾矿中的金属硫化物与氧化剂(如大气中的氧气和高价铁)接触,进而降低尾矿中金属硫化物的氧化率。

3.4 基因工程

借助先进的基因工程工具,可以设计具有所需特性的微生物,它们具有螯合金属的蛋白质和多肽,并积累重金属。受重金属污染场所生长的微生物通过调节多种遗传机制适应环境重金属的毒性。这些菌株具有在污染场所生存的能力,通过从此类菌株引入单个基因或基因簇能够培育出强修复能力的重组菌株(Gadd et al., 2005)。具有强修复能力的菌株具有两个特征:耐受性强,即在高浓度重金属环境下能够生存;能够通过某种机制有效的去除重金属。最常见的技术包括对单个基因或操纵子进行工程改造、现有基因的基因序列改变以及途径转换(表7),如通过基因改造的大肠杆菌JM109、蜡状芽孢杆菌BW-03(pPW-05)和金属利尿铜霉菌株MSR33能够有效的修复Hg污染(Ruiz et al., 2011; Rojas et

al., 2011; Dash et al., 2014; Dash and Das, 2015),基因改造的铜绿假单胞菌和叶绿素微藻菌株则具有较强的Cd耐受性和修复潜力(Kermani et al., 2010; Ibuota et al., 2017)。当然,在某些情况下单基因操作不能与天然适应菌株竞争,如基因改造的叶绿素微藻菌株具有一定的Cd耐受性和修复能力,但是受氧化应激耐受性有限的影响,其效果并不比天然适应菌株优越(Ibuota et al., 2017)。

4 动物修复技术

动物修复技术是利用土壤中某些动物吸收土壤中的重金属,降低污染土壤中重金属的含量。动物通常没有明显的代谢重金属的能力,无法将其还原为完全无毒或毒性较小的状态。少量研究表明部分动物能够富集重金属从而有效降低环境中重金属的含量,其中研究程度最高的为蚯蚓,蚯蚓在重金属污染修复中有2个作用:一是可以通过自身的吸收来富集重金属,从而降低土壤重金属含量,如蚯蚓的黄色细胞对Cd和Pb有较强的吸收力,可以利用体内的金属硫蛋白固定金属,生成无生物毒性的镉—金属硫蛋白、铅—金属硫蛋白形态,从而富集重金属,实现修复功能(Ireland, 1983);二是可以通过自身的活动改善土壤中重金属的活化能力,从而促进植物对其富集,如蚯蚓在土壤中活动能够分泌出大量

表6 实验室生物淋滤金特征统计

Table 6 Characteristics of bioleaching gold in laboratory

微生物	金属来源	Au采收率	机制	pH	温度(℃)	资料来源
巨大芽孢杆菌(<i>Bacillus megaterium</i>),铜绿假单胞菌(<i>Pseudomonas aeruginosa</i>)	纯金颗粒和氧化的含金矿石	超过80%	产氰	10	35	Gorji et al., 2020
粪曲霉(<i>Alcaligenes faecalis</i>)	尾矿	18.5%	产氰			Pineda, 2019
嗜酸氧化亚铁硫杆菌(Acidithiobacillus ferrooxidans),硫氢酸性硫杆菌(Acidithiobacillus thiooxidans),钩端螺旋体杆菌(<i>Leptospirillum ferrooxidans</i>)	多金属硫化物矿	90% (Ag=80%)	含硫配体	初始为2	30	Kržanovic et al., 2019
嗜压玫瑰变色菌(<i>Roseovarius tolerans</i>)	Modi Taung金矿	100%	含碘配体	7.7~8.4	30~35	Khaing et al., 2019
酸性硫氧化铁杆菌(<i>Acidithiobacillus ferrooxidans</i>)	难处理金矿	85%	含硫配体	1.75	32	de Carvalho et al., 2019
嗜酸细菌	含金碳质、泥质板岩、黄铁矿原料	92%和88%	含硫配体		25	Daibova et al., 2019

表 7 细菌基因工程提高有毒金属的生物修复潜力

Table 7 Genetic engineering in bacteria to enhance heavy metal bioremediation potential

细菌	遗传机制	遗传改变	应用	资料来源
大肠杆菌 (<i>Escherichia coli</i>) JM109	质粒	引入 mt-1、ppk 基因	高效的 Hg 积累/转化	Ruiz et al. , 2011
蜡状芽孢杆菌 (<i>Bacillus thuringiensis</i>) BW-03 (pPW-05)	MerA	引入 merA 基因	有效的 Hg 挥发和生物吸附	Dash and Das, 2015; Dash et al. , 2014
金属利尿铜霉 (<i>Cupriavidus metallidurans</i>) MSR33	Mer 操纵子	merB, merG 和其他 mer 基因的改变	对无机和有机 Hg 以及 Cu 和 C 的耐受性	Rojas et al. , 2011
铜绿假单胞菌 (<i>Pseudomonas aeruginosa</i>)	Cad 操纵子	cr 喹橙和 a 喹黄素对 cad 操纵子的突变	增强 Cd ²⁺ 耐受性	Kermani et al. , 2010
大肠杆菌 (<i>Escherichia coli</i>)	质粒	食芳烃新鞘氨醇菌的 NiCoT 和 E. coli 的合成黏附操纵子	增加 Ni ²⁺ 和 Co ²⁺ 封存以用于生物过滤	Duprey et al. , 2014
谷氨酸棒杆菌 (<i>Corynebacterium glutamicum</i>)	Ars 操纵子 (ars1 和 ars2)	放线菌 Mrx-1	增强 As 耐受性并将 As ⁵⁺ 还原为 As ³⁺	Mateos et al. , 2017
叶绿素微藻菌株 (<i>Chlorella luteoviridis</i> , <i>Parachlorella hussii</i> , <i>Parachlorella kessleri</i>)	CrMTP4 基因	CrMTP4 基因过表达	增强 Cd 耐受性以及修复潜力	Ibuota et al. , 2017

成分复杂的胶黏物质,其中很多成分含有一COOH 和—C=O,它们能够络合、螯合重金属,提高土壤中重金属的活性,但是这种作用的速率很慢,需要更多的研究提高其性能(Wu Gang et al. , 2010),且蚯蚓对不同金属的活化程度也不同,如赤子爱胜蚓 (*Eisenia fetida*)能够增强植物蚕豆和玉米对 Cd 的吸收(Lemtiri et al. , 2016);此外,蚯蚓粪是很好的重金属修复剂,具有很好的通气性、排水性和高持水性,可增加土壤孔隙度和团聚体数量,同时具有很大的表面积和较强的吸附能力,可在很大程度上吸附重金属(Singh et al. , 2020; 李扬等, 2010)。

对蚯蚓修复重金属污染的影响因素使多种多样的,包括蚯蚓类型、土壤性质、气候以及植被发育情况等。安德爱胜蚓 (*Eisenia andrei*)在低 pH 和温度下对土壤 Hg 具有更强的生物积累能力(Le Roux et al. , 2016);在增强蚯蚓修复能力方面,土壤中加入全氟烷基酸(PFAAs)能够增强蚯蚓对 Cd、Zn、Ni、Pb 和 Cu 的吸收能力(Zhao Shuyan et al. , 2018);同一生理生态群不同种的蚯蚓对金属污染物的修复具有相似的特性,均对重金属有较高的富集能力(Wang Kun et al. , 2018);但它们对不同金属的富集能力具有较大的差异,其生物累积因子分别为: Cd(10.6 ~ 18.8)、Zn(1.15 ~ 1.75)、Cu(1.01 ~ 1.35)、Pb(0.56 ~ 0.95)(Wang Kun et al. , 2018)。

5 结论和展望

通过对物理化学、植物修复、微生物修复和动物修复 4 大类矿山重金属污染修复技术现状和进展进行详细梳理,初步结论如下:

(1) 物理修复技术相对简单有效,但从修复时效角度考虑存在失效的风险,因此基本不能满足当前环境修复的需求。

(2) 化学修复技术是目前常用的方法之一,包括化学固定、化学浸出、化学转化 3 种机理,其中化学固定和化学浸出的研究程度更高、应用范围更广。但是,化学修复需要消耗大量化学试剂和新材料,因此修复成本较高。化学修复重点研究方向包括:高效、绿色、可重复利用新型修复材料(如纳米材料)研发和常规廉价材料的优化应用(如粉煤灰、石灰、腐殖质、堆肥)。

(3) 植物修复技术作为绿色、廉价和有效的修复途径,在金属矿山重金属污染土壤修复领域拥有巨大优势和潜力。但是植物修复面临着修复周期长、修复不彻底以及修复后受污染生物质处理等问题。未来的研究方面主要集中在两个方面:基因工程,使得能够增强修复能力的基因传导到生物中并得到相应的表达,增强生物的重金属耐受性和修复能力(如积累或转化),最终缩短修复周期、提高修

复效率;针对经济价值较高的重金属(Au、Ni和Cu等)污染治理,可以尝试采用植物采矿技术进行修复,利用超积累植物修复重金属污染的同时回收这些金属,从而大幅降低综合成本。

(4)微生物修复技术,作为一种绿色、高效的修复途径,在金属矿山重金属修复领域前景十分广阔。但是,现阶段微生物修复技术主要集中在实验室和小规模现场试验阶段,实际工程应用有限。需要通过基因工程等生物技术增强微生物对环境的耐受性和修复效率。

(5)动物修复技术的研究则主要是针对蚯蚓,这种技术应用十分有限,更多的是辅助植物修复技术增强修复效果。

此外,为了更好的解决重金属污染问题,要促进不同方法技术的综合应用,取长补短,以实现更好的修复效果;加强不同修复技术的数据库和智能决策系统建设,针对不同重金属污染类型、程度、环境条件及修复效果能够给出最优解决方案和快速定量评价;同时,可以转变思维,基于不同植物对重金属富集效应不同以及重金属在植物不同部位富集的原理,选取对特定重金属不具富集效应、且在果实部位富集效应最弱的植物进行种植,修复生态环境的同时实现经济效益且避免重金属进入人体。

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Progress and prospect of remediation technology of heavy-metal-contaminated soil in mines

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Abstract: As more and more attention has been paid to the remediation of heavy-metal-contaminated soil in mines, researchers have achieved considerable progress in various remediation technologies, and new remediation technologies play an increasingly important role in the remediation of heavy-metal-contaminated soil in mines. This paper has made a detailed review of theoretical, experimental, and practical progress of remediation technologies for heavy metal contaminated soil, including physical and chemical remediation, phytoremediation, microorganism-based remediation, and animal-based remediation. Then, it is summarized and proposed that the remediation technology of heavy metal contaminated soil in mines will focus on three directions: low-cost-oriented conventional technology optimization, including the effective use of conventional cheap materials, and recovery of the targeted heavy metals (hyperaccumulator and chemical recovery); technology-oriented efficiency improvement, including the research and development of new and efficient remediation materials such as nanomaterials and biofilms, genetic engineering and other top technologies, greatly improve remediation efficiency and reduce overall costs through the fine research of micro-mechanisms of each remediation technologies; combined remediation technologies, such as microbial—plant, chemical—plant, physical—chemical and other combined remediation technologies, achieve better remediation results by integrating the advantages of each remediation technologies; strengthen the construction of databases and intelligent decision-making systems for different restoration targets to promote the commercialization of research findings.

Keywords: Heavy-metal-contaminated soil; remediation technology; phytoremediation; microbial-mediated remediation; nanomaterial

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